

REVIEW OF OPTIONS FOR MANAGING WASTE FROM THE HUNTERSTON A SOLID ACTIVE WASTE BUNKERS GIVEN ITS STATUS AT MAY 2021

WD/REP/0032/21 Issue 2

by **by** Renior Waste Consultant, Programmes Strategy, Magnox Limited

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EXECUTIVE SUMMARY

Scope of Assessment

This assessment determines the management strategy for waste arising from the solid active waste bunkers ('the bunkers') at Hunterston A site. It is a strategic options assessment which forms part of the demonstration that Best Practicable Means (BPM) have been applied to manage the waste in scope.

The waste in scope is that identified in the radioactive waste inventory under waste streams 9J19 to 9J23 inclusive, arising from all five bunkers. Waste from Bunker 1 consists primarily of magnox fuel element debris (FED) and waste from Bunkers 2-5 consists primarily of graphite FED, though there are smaller quantities of miscellaneous activated components and contaminated items intimately mixed with these wastes along with sludge and graphite particulate contamination. Segregable fuel fragments were excluded from assessment as these are being considered by a separate study.

The starting point of this assessment takes account of the status of the waste as at May 2021, the majority of which had been retrieved and containerised into 3m³ boxes (1,000 of ca. 1,100 boxes filled with only waste from Bunker 1 remaining to be retrieved).

The extant (baseline) strategy, supported by a 2012 strategic options assessment [1], is to encapsulate all waste for subsequent storage on-site until a final management route exists, with disposal in a future near-surface facility being the envisaged end point.

This new assessment was initiated in response to the Scottish Environment Protection Agency's (SEPA) challenges to the extant strategy. In November 2020 Magnox Ltd committed to review the BPM case and its supporting characterisation.

Method of Assessment

Existing characterisation data were reviewed prior to conducting the strategic options assessment to ensure that the data quality was appropriate for this exercise. This review concluded that the existing waste inventories had been developed to support the Letter of Compliance (LoC) and were appropriate for this purpose but included conservatisms. It was decided to produce new inventories to support the strategic options assessment so that potential management options were not obscured by the conservatisms of the existing inventories.

Credible options were then defined based on analyses of these revised inventories. This was achieved by considering the feasibility of treatment and disposal options for materials present within the waste. These analyses used existing waste acceptance criteria (WAC) where available, for example that of the Low Level Waste (LLW) Repository, else relied on best available information, for example the joint environment agencies' Guidance on the Requirements for Authorisation (GRA) of near-surface disposal facilities. These analyses concluded that the following credible options should be assessed:

The treatment options shown above seek to reduce the volume of higher activity waste (HAW) requiring long-term management, of which only dissolution is commercially available today. No material recovery options were identified. Two forms of near-surface disposal were considered:

- At-depth disposal where waste packages are disposed of in vaults or caverns which are \bullet constructed several tens of metres below the surface and accessed through a drift. No such facility exists currently, though it is planned to be implemented under Scotland's HAW Policy Implementation Strategy [2].
- At surface disposal where waste packages are disposed of in vaults which are constructed \bullet at or close to the surface and backfilled when full (the protective covering is of the order of a few metres thick); this type of facility is typically for waste categorised as LLW. The LLW Repository is such a facility.

All alternative options to at-depth disposal require the waste to undergo further processing, e.g. sorting and segregation. A short-list of options was developed according to whether this processing takes place promptly or is deferred until the reactor decommissioning phase:

In accordance with Company Standard S-391 [3], these options were assessed using an attribute comparison method to determine a 'lead option', by considering the relative strengths and weaknesses of each option within each attribute, then risks were considered in a 'management analysis' to identify a 'proposed option'.

This assessment separately considered options for managing waste from Bunkers 2-5 and Bunker 1. For waste from Bunker 1 this assessment also separately considered the waste remaining in bunker and that which had already been retrieved and containerised; this is because the waste remaining in bunker has a greater potential to be alternatively managed. For waste that has already been retrieved and containerised into 3m³ boxes (this applies to all waste from Bunkers 2-5 and roughly half of the waste from Bunker 1), all alternative options to at-depth disposal require the waste to be retrieved from these boxes for processing.

¹ Wherever possible this assessment has considered definite end points, however 'long-term storage' has been applied where no credible end points could be determined. This envisages up to 300 years of storage whilst final management solutions are developed. This assessment has assumed that wastes remaining after processing or treatment are unsuitable for at-depth disposal or any other end point that can be defined at present within the constraints of Scotland's HAW Policy [6].

Outcome of Assessment

The proposed option is to manage all waste within the scope of this review by encapsulation for interim storage then at-depth disposal (Options G1 and M1). This aligns with the baseline strategy².

Treatment options were judged to perform less well than disposal options because:

- All treatment options require new facilities to process the waste, requiring additional waste handling and resulting in increases to worker dose. There would also be conventional safety risks from the construction, operation and decommissioning of these facilities, and particularly for the scenarios where waste must first be retrieved from $3m³$ boxes.
- Most treatment options apply inherently high hazard processes and, aside from dissolution, all technologies have low technology readiness levels and credible development risks.
- All treatment options would result in gaseous and/or liquid wastes to manage. Even following abatement it was estimated that doses due to radioactive discharges would be two or more orders of magnitude greater than those with the disposal (encapsulation) options. The estimated total doses (following abatement) due to discharges for each option 3 are shown below in µSv/y:

Applying BPM, as set out in SEPA guidance [4], presents an imperative to minimise radioactive waste discharged to the environment, a necessary consequence of which is that the amount sent to solid disposal facilities is maximised.

- All treatment options introduce technical challenges to adequately sort and segregate the waste to the requirements of the relevant process.
- All treatment options result in HAW which is unlikely to be suitable for at-depth disposal.
- Some treatment options would incur significantly greater cost. The estimated total cost of each option³ is shown below:

Alternative disposal options were only identified for magnox from Bunker 1, however at-surface disposal performed less well than at-depth disposal because:

- The magnox would need to be segregated from the other waste from Bunker 1 and there are risks that a portion of the magnox cannot be adequately sorted to meet at-surface disposal WAC. Challenges are presented by the heterogeneity of the waste and presence of small, high activity items, as well as particulate contamination.
- Sorting the magnox and exporting it requires additional waste handling and results in increased worker dose due to this.
- For the waste remaining in bunker the requirement for a new sorting facility would present competition for resources at site and would impact the site schedule. This issue is significantly exacerbated for the waste that has already been retrieved and containerised.
- There is greater uncertainty with management of the segregated, non-magnox waste as this is unlikely to be suitable for at-depth disposal if packaged on its own.
- At-surface disposal would also incur additional waste-miles and cost.

² The baseline strategy is encapsulation for long-term storage until a final management solution exists, with disposal envisaged as the end point. The proposed option aligns with this though is clearer about the envisaged end point.

³ Figures presented for Bunker 1 options (M1-6) are for the 'in bunker' portion of the waste only.

The proposed option, to manage all waste by encapsulation for storage then at-depth disposal, follows an underpinned waste management approach and is assured via the LoC process. The key risks with this approach relate to disposability and foreclosure of options. The possibility to defer encapsulation has been considered as one tactic to mitigate these risks, however this review has identified no grounds for pursuing such an approach. Prompt encapsulation is proposed to ensure that the waste packages are manufactured in accordance with the design intent of the $3m³$ box, to provide better long-term performance during storage and during the disposal facility's operational and post closure phases. During storage the grout limits waste and container degradation, assuring their ongoing performance; following disposal the grout provides retardation of release of radionuclides into the biosphere and stabilises chemically reactive or mobile materials, as well as providing a structural function.

Implementation of this strategy will continue to be managed using the Hunterston A site Radioactive Waste Management Case, which will be updated to reflect the findings of this review.

BPM Summary Statement

This document presents a strategic options assessment. The proposed option is considered to represent BPM and supports the requirement to optimise public exposures as:

- it results in minimal generation of further radioactive wastes (secondary wastes are limited to operational and decommissioning wastes associated with existing retrieval and encapsulation processes/equipment, which has arisen/will arise in any case); and
- it minimises the potential for radioactivity to be discharged to the environment (discharges would arise primarily from encapsulation off-gassing, which would be similar in all disposal options assessed; treatment options would result in comparatively significant discharges).

The strategic assessment presented in this document has not addressed how the implementation of the proposed option will be optimised. To clarify how this is being achieved an optimisation summary report will be produced. This will set out how the radiological effects of radiological discharges on the environment and people have been or will be minimised.

ALARP Summary Statement

The proposed option is considered to represent ALARP. No additional handling of the waste is proposed (whereas all other options would involve this) and it makes use of the site's established HAW management route which is remotely operated by and large. There would be minimal off-site doses due to discharges compared with treatment options. There are other options which would result in fewer radwaste consignments, and potentially lower public doses due to transports, though these are overall less preferable due to other factors. Long-term risks are managed through prompt waste passivation and applying an underpinned waste management approach.

Forward Actions

No further work is proposed to support the outcome of this strategic options assessment. Forward actions may be needed to optimise its implementation, these will be identified upon completion of an optimisation summary report.

Whilst not forming a forward action from this review, it is noted that work is ongoing to support the NDA in developing a NSD capability and that this work, along with collaborations with other Scottish waste producers, will hopefully improve how disposal uncertainties are factored into near-term decisions.

Review Date

This is to be periodically reviewed every five years in accordance with company standard S-391, or sooner if prompted by a trigger (see below), until the waste is encapsulated. Following encapsulation this should only be reviewed if prompted by one of the triggers.

Review Triggers

This should be reviewed if:

- LLWR WAC change to enable disposal of the waste as packaged.
- Other alternative management routes become available and which challenge the assumptions used in this review.

* Verification completed by Michelle Grist prior to leaving the company. Evidence of this verification has been confirmed by Elena Alcantara and minor amendments made since verification have been confirmed as appropriate.

SEND A COPY OF THE FULLY SIGNED REPORT TO ANY OPTIONS ASSESSMENT **OVERSIGHT MANAGER.**

⁴ After consulting all functions as appropriate (Asset Management Programme Manager; Decommissioning Director; EHSS&Q Director; Waste Strategy & Permissioning Manager) and affected sites as necessary. Completed Form is a non-permanent record Page 8 of 81

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1 INTRODUCTION

1.1 Background

The Hunterston A Solid Active Waste Building (SAWB) is a large, two storey reinforced concrete building adjoining the south west corner of the Cartridge Cooling Pond facility. The SAWB was designed principally for the storage of solid intermediate level waste (ILW), comprising five large contiguous reinforced 'bunkers' for this purpose. Bunkers 1-3 were constructed in 1964 with Bunkers 4-5 having been added in the 1980s.

The waste stored in the SAWB arose from the removal of fuel furniture and associated fuel channel components, comprising both the primary materials and operational waste arising from these activities. Upon discharge from the reactors, the irradiated fuel assemblies were processed to separate the graphite sleeve from the fuel element; the graphite was primarily discharged to Bunkers 2 to 5. The element was then desplittered to separate the magnox splitters and fuel cans, and the waste magnox splitters were primarily sent to Bunker 1. A number of other waste items were also discharged to the bunkers and these became intimately mixed with the graphite and magnox fuel element debris (FED) (see Section 1.3 and 2.1.2).

The baseline strategy is to retrieve and place these mixed wastes into thin-walled, stainless steel containers (' $3m³$ boxes') for encapsulation then storage on-site until a final management route becomes available. The basis for this strategy is set out in Reference [1], as amended by Reference [5].

Retrievals commenced in 2014 with the bunkers being accessed sequentially from Bunker 5 to Bunker 1. All the waste from Bunkers $2 - 5$ has been retrieved and Bunker 1 retrievals are in progress. All waste retrieved has been 'containerised' and is planned to be encapsulated in the nearterm.

The site Lifetime Plan (LTP) allows these encapsulated waste packages to be stored for up to 300 years whilst management options are developed in accordance with Scotland's Higher Activity Waste (HAW) Policy – facilities for the storage, treatment and/or disposal of HAW will be developed as part of the Policy's implementation. The Policy is that such facilities will be as near to the site where the waste is produced as practicable, and that the facilities will be in the near-surface environment [6]. In contrast to policy in England and Wales, Scotland's HAW Policy prevents direct disposal to the Geological Disposal Facility (GDF).

1.1.1 Requirement for Review

In late 2019 an application was made to vary the site's environmental permit to include a new authorised gaseous outlet for SILWE. As part of their assessment, the Scottish Environment Protection Agency (SEPA) presented Magnox with challenges to the SAWB waste strategy and Magnox committed to reviewing the BPM case and its supporting characterisation.

The history and basis for the extant strategy have been explored in previous engagements between Magnox and SEPA. This review is not required to address such history, only to consider the correct course of action given the current situation.

1.2 **What is the Decision to Be Made?**

What is the management strategy for the bunkers waste?

This review should identify the optimal management strategy and determine the near-term actions, such as whether to segregate the wastes or not, or whether to encapsulate the waste or not.

The conclusions may be specific to waste originating from Bunker 1 and waste originating from Bunkers $2 - 5$, as the waste compositions differ between these subsets. The conclusions may also be specific to waste that is already retrieved and that which is yet to be retrieved.

13 **Scope**

The scope of this review is to identify the preferred strategy for managing waste from the SAWB bunkers ('the bunkers').

The scope of this review applies to waste defined within the radioactive waste inventory (RWI) waste streams 9J19–9J23. The waste types/materials present within these streams can be grouped as follows⁵ [7], arranged by their percentage of total inventory (by volume):

- Graphite (fuel element sleeves) (67.6%) (mainly present in Bunkers 2-5) \blacksquare
- Magnox (fuel element splitters) (25.5%) (mainly present in Bunker 1) \blacksquare
- Miscellaneous Contaminated Items (MCI) (filter dust bags, filters, incinerator waste, general \bullet waste, pond sludge, pond skips) (2.9%)
- Mild steel / iron (fuel support members (FSMs) and thermocouple reeling drums) (2.3%) \blacksquare
- Zirconium (D-bars) (1.7%) \bullet
- Stainless steel (thermocouples, burst cartridge detector (BCD) clips, contact assemblies, \blacksquare control rod wires, F.P. wires and bobs) (0.07%)
- Aluminium Chromium (thermocouple cables) (0.004%)

These waste items were stored together in the bunkers and became intimately mixed. In addition to the pond sludge there is particulate waste in the form of graphite dust.

The waste can broadly be described within two subsets: waste from Bunkers $2 - 5$ and waste from **Bunker 1:**

Table 1: Summary of waste inventory

⁵ N.B. Uranium is omitted from this breakdown (segregable uranium is excluded from the scope of this review) though, for information, it comprises <0.01% of the total inventory.

The starting point of this review takes account of the status of the waste at May 2021, where 890 boxes had been filled with waste from Bunkers $2 - 5$ (these bunkers are now empty) and 110 boxes had been filled with waste from Bunker 1 (with an expected 100 boxes remaining to fill). All waste retrieved has been containerised into 3m³ boxes in accordance with existing Letter of Compliance (LoC) endorsements⁶. Figure 1 illustrates the bunkers and how waste has been / is being retrieved from them: a pile of mixed waste is shown on the floor of an emptied bunker, having been pulled through from the adjacent bunker, where it is sorted and loaded into metallic buckets for extraction to a separate location for packaging. The extent of graphite dust present in the waste is illustrated by the floor where the remotely operated vehicle has traversed.

This review considers all options for managing these wastes throughout the remaining lifecycle such that any (residual) HAW could be safely stored for up to 300 years. This includes options which would reduce the amount of HAW requiring storage throughout this period (or a portion of it). Consideration is also given to the risk of foreclosing options and, more broadly, alignment with Scotland's HAW Policy.

This review considers which management options are available from this point forwards and what would be required to enable them. This review does not account for historical actions. As the site is set up to deliver the baseline option this therefore presents an inherent advantage for it and, whilst all options are assessed fairly given their respective statuses, this is acknowledged.

An encapsulation plant (the Solid ILW Encapsulation plant, SILWE) already exists on-site and is entering its commissioning phase, with active commissioning due to commence in mid-2023. A waste package storage facility (the ILW Store) already exists on-site and is operational.

⁶ To provide assurance that *the future management of waste packages has been taken into account as an integral part of their development and manufacture* [2]*,* and in accordance with regulator expectations, the LoC process has been followed during implementation of the project. Nuclear Waste Services (NWS) has issued Magnox with final stage LoCs (fLoCs) for the containerisation and encapsulation of waste from all bunkers.

1.3.1 Exclusions

A small volume, particulate waste stream (9J63), arising from processing the bunkers waste, is excluded from this review. No alternative management options have been identified for this waste.

Segregable⁷ fuel debris is excluded from assessment. A separate study has been commissioned to determine how these should be managed.

1.4 Previous or Related Studies

Two previous related studies are noted that support the case to retrieve, package and encapsulate the waste, followed by long-term storage in accordance with Scotland's HAW Policy.

- Hunterston A Solid Active Waste Buildings Bunkers Waste (9J18–9J30 / 9J35–9J42 / 9J44) Best Practicable Environmental Option (BPEO) / Strategic Options Assessment, HNA/2981/PJ/SR/1131 Issue 1, November 2012 [1].
- LC35 Change Justification, Scottish Sites Long-Term Storage Strategy, M/WF/GEN/REP/0003/15, Issue 1, September 2015 [5].

1.5 Engagement

An options assessment panel has been used for this study, the details of which, along with other engagements, are included in Appendix A.

1.6 Characterisation Data Quality

A characterisation study [8] has been conducted to determine whether the existing data are suitable for this review.

This study determined the basis of the current inventories, which had been developed in support of LoC submissions and were principally based on neutron activation modelling, supplemented by measurement of magnox and pond sludge samples. This approach was deemed to be fit-for-purpose in respect of the LoC but based on conservative assumptions, principally the use of upper bound element precursor concentrations.

To ensure that the data did not obscure any waste management options for consideration it was recommended that the activation modelling be re-performed using best estimate element precursors and utilising modern modelling tools. Additionally, it was proposed that the data be further improved by using bunker filling histories to refine the decay times and irradiations experienced by the different waste materials deposited at different times in the bunkers. Of note, existing in-depth reviews of high dose rate items (activated components including thermocouples, FSMs, and D-Bars) could be consulted to provide accurate information on material compositions and deposition histories. The recommendations from this study have since been enacted. This has involved further characterisation through scrutiny of existing information and neutron activation modelling. The revised inventories, documented in Reference [9], have been used as the basis for this review.

As part of this review the option to take further sample measurements to support this re-modelling was evaluated but decided against. This was because the remaining waste being retrieved is principally magnox, which has already been sampled, and taking samples from containerised waste was viewed to be grossly disproportionate to the benefit gained for this review. Aside from ALARP and cost considerations, amongst others, retrieving samples from containerised waste is problematic from a statistical perspective; any data obtained are unlikely to be representative of equivalent components within the population as a whole, as these have experienced difference irradiation and decay histories. Neutron activation modelling, however, was viewed to provide data of sufficient quality to support this review.

 $\frac{7}{1}$ This review assumes the definition of segregable aligns with the installed fuel detection system which is configured to identify items of fuel debris greater than 380g during retrieval of waste from Bunker 1.

OPTIONS $\mathbf{2}$

Derivation of the Options for Assessment 2.1

2.1.1 Long-list of Options

The management options for the waste can be classified into three groups and summarised as shown below, where the waste is either managed in 3m³ containers per the current strategy or undergoes further processing, either now or in the future, to enable alternative management. The options for alternative management seek greater application of the waste hierarchy and include variations of treatment⁸ and disposal.

Alternative management options can be grouped into two categories, prompt processing and deferred processing, where 'processing' refers to additional work performed on-site to enable alternative management, such as further waste sorting and/or packaging. The benefit of prompt processing is that the site is currently mobilised and could perform such work. The benefit of deferred processing is that it allows time for treatment or disposal routes to become available before processing is undertaken.

Table 2: Grouping of options within the long list

Notes:

Scotland's HAW Policy requires facilities for the management of HAW to be in the near-surface environment; for disposal facilities this is referred to as near surface disposal (NSD). Two variations of NSD have been considered when generating the long list of options (see Appendix B for more information):

- At-depth disposal where waste packages are disposed of in vaults or caverns which are \bullet constructed several tens of metres below the surface and accessed through a drift.
- At surface disposal where waste packages are disposed of in vaults which are constructed at or close to the surface and backfilled when full (the protective covering is of

⁸ For example, treatment to enable reuse or recycling of materials, or to achieve volume reductions to minimise subsequent demands on storage capacity etc.

the order of a few metres thick); this type of facility is typically for waste categorised as Low Level Waste (LLW).

For wastes that cannot be disposed of in the near-surface environment, and which cannot be otherwise managed by treatment, long-term storage is the only currently available management option⁹. The Magnox LTP allows for packaged HAW to be stored long-term, for up to 300 years, whilst final management solutions are developed.

2.1.2 Option Screening

The principal screening criterion is that deferred processing is sub-optimal for options that are currently feasible.

To further screen these options the following constraints have been identified:

- All waste that has been retrieved already is containerised within 3m³ boxes and stored within the ILW Store. Alternative management options which require the waste to undergo further processing will therefore require these containers to first be emptied.
- A deep geological disposal option, such as the GDF adopted in England and Wales, is not available within Scotland's HAW Policy. It will be considered in the management analysis as an opportunity, should Policy change to permit it. This is explored in Section 6.
- Encapsulation is required for all disposal options, and deferral of encapsulation is not considered as a distinct option. This is explored in Section 6.
- At-surface disposal is not credible for unsorted waste based on current criteria (see Appendix F).

The following key assumptions are also made:

- It is feasible to sort and segregate the waste into its constituent materials, to enable their separate management.
- It is feasible to dispose (NSD) of the waste as currently packaged.
- It is not feasible to dispose (NSD) of waste remaining following processing or treatment.

A full list of constraints and assumptions is included in Appendix B including a reasoning for each assumption.

2.1.2.1 Applicability of options to certain materials

The management options potentially applicable to each of the main materials present in the bunkers waste are shown in the following table. This is supported by:

- Reference [10], which reviews the feasibility of at-depth disposal using the Guidance on Requirements for Authorisation (GRA) [11] as its basis. See Appendix G.
- Reference [12] which reviews the feasibility of at-surface disposal using the current Low Level Waste Repository (LLWR) Waste Acceptance Criteria (WAC) as its basis. See Appendix F.
- Appendix B which reviews the feasibility of treatment options.
- Assumptions in Appendix B.

⁹ Wherever possible this assessment has considered definite end points, however 'long-term storage' has been applied where no credible end points could be determined. Indefinite storage is not viewed as a strategy in its own right.

Table 3: Option Applicability by Material Type

As shown in Table 3, no viable alternative options were identified for managing mild steel. MCI. stainless steel, aluminium, and chromium. The remainder of this assessment therefore focusses on options for managing magnox and graphite, for which there are alternatives.

2.1.3 Short-list of Options

Based on the split of waste materials between the two bunker subsets. Bunkers 2-5 and Bunker 1, the following can be asserted:

- alternative management options for graphite only apply to waste from Bunkers 2-5, and
- alternative management options for magnox only apply to waste from Bunker 1. \bullet

This is because less than 1% of graphite is/was stored in Bunker 1, and less than 1% of magnox was stored in Bunkers 2-5. The detriment of retrieving these small quantities from either subset is viewed as grossly disproportionate to any benefit (additionally, it is not recorded whether any single package contains either of these materials, therefore targeted retrievals would not be possible for any waste already containerised). Accordingly, management of the other waste materials (mild steel. zirconium. MCI, stainless steel, aluminium chromium) will also be considered within these subsets.

The options can therefore be presented according to the management strategy for the dominant waste material within each bunker subset. The short-listed options are summarised in the Table 4 and illustrated in Figure 2 (these have been assigned numbers for ease of discussion in the following sections). A description of these options is provided in Appendix B. A break-down of the steps involved with each of these options is given in Appendix E, which shows how all waste materials from Bunkers 2-5 and Bunker 1 will be managed, as well as any waste remaining following processing or treatment. The assessment summary within the main body of this document is presented separately for Bunkers 2-5 and Bunker 1; see Section 4 and Section 5 respectively.

Table 4: Shortlisted Options

¹⁰ Option of prompt processing for gasification eliminated as sub-optimal to deferred processing (see Appendix B). Completed Form is a non-permanent record Page 17 of 81

ASSESSMENT METHODOLOGY 3

3.1 **Method**

In accordance with Company Standard S-391 [3], the lead option for managing waste from each bunker subset will be determined by comparing the attributes of each option and using reasoned arguments, and the proposed option will be determined following a management analysis of the lead option.

3.2 **Attributes**

The selected attributes used in the assessment are shown below and Appendix C presents the rationale for any that are excluded.

3.3 **Assessment Data**

See Appendix D.

DETERMINATION OF THE LEAD OPTION FOR BUNKERS 2-5 4

4.1 **Bunkers 2-5 Waste Summary**

The inventory of waste from Bunkers 2-5 is summarised below. All waste has been retrieved and is in storage in 890 3m³ boxes in the ILW Store. During packaging the waste was processed (sorted, size reduced, etc.) according to the requirements of the LoC Waste Product Specification (WPrS) [13] which includes limits on certain waste items, such as fuel support members and HEPA filters, and identifies measures to segregate particulate waste.

Summary of Bunkers 2-5 inventory (9J19-22)			
Total raw volume	1660 m^3 (approx. 2,180 tonnes of raw waste)		
Waste type	Mainly graphite and metallic debris, with only small quantities of misc ILW; there are also small amounts of magnox in Bunkers 2-3.		
Dominant radionuclides	Ni-63, H-3, Co-60, C-14	Graphite MCl ■ Zirconium Aluminium chromium	■ Magnox Mild steel / iron Stainless steel

 $\overline{42}$ Assessment of Options for Managing Waste from Bunkers 2-5

This section presents a comparison of options based on the attribute comparison table in Appendix E and the assumed steps involved with each of the options set out in Appendix B.

The attributes of each option have been assessed and 'scored' relative to each other by virtue of assigning a green/amber/red colour according to the system set out in Reference [3], where:

- **Red is allocated where the performance of the option on that issue is significantly worse than** a. at least one other option.
- Amber is allocated where the options is clearly, but not significantly, worse than at least one \mathbf{r} other option.
- Green is allocated in other cases. ×

The relative performance of each option is discussed within each attribute category as follows:

4.2.1 Safety Attributes

The RAG 'score' for each option against each attribute (from Appendix E) is shown below:

The key differences between these options are:

- Option G1 makes use of existing plant and processes without modification whereas Option G2 requires a new 3m³ box emptying facility and processes to sort, assay and export graphite in 200 drums and package the remaining waste into 3m³ boxes. These extra requirements for Option G2 present conventional safety risks due to construction and implementation. Considering the processes used with each option. Option G1 requires the encapsulation of 890 boxes with associated hazards, e.g. due to the use of cement. Option G2 requires far fewer boxes to be encapsulated (estimated at 132) but presents inherent hazards due to the gasification process which involves graphite crushing and a plasma furnace. On balance, and along with considering risks posed by its novelty. Option G2 is viewed to present the overall greater safety risk due to these inherent process hazards.
- In Option G1 the waste would be promptly passivated by encapsulation, minimising risks \mathbf{r} during the storage phase. In Option G2 the graphite would not be treated until the reactor decommissioning phase and so presents a comparatively long time at increased risk due to 'raw' storage (though see Section 6.3.1 for consideration of these risks).
- Greater worker doses would be expected with Option G2 because the waste would \bullet undergo further handling and, whilst the work would be managed in such a way as to ensure doses are ALARP, the overall ALARP option would be one which avoids the work in the first place, i.e. Option G1.
- Option G1 incurs the most radioactive transports (890) compared with Option G2 (323). \bullet Public doses are considered to be greater for the former, however this assessment has not considered the shielding properties of the different transport packages. Conventional traffic risks would be greater for Option G2 due to an increase in waste-miles and the need to import additional materials to site for construction.

In summary, Option G1 performs best on safety as it makes use of existing plant and processes. and the waste is promptly passivated (by encapsulation). Option G2 performs less well as it involves new facilities and processes being established at site, increased handling of the waste, an inherently high hazard process, and delays to passivation.

4.2.2 Environment, Waste and Sustainability Attributes

The RAG 'score' for each option against each attribute (from Appendix E) is shown below:

The key differences between these options are:

Option G1 would dispose of all waste without first reducing its volume or activity whereas \bullet Option G2 would result in a greatly reduced volume of waste being disposed, the activity of which would be reduced (due to discharges during treatment). Option G2 would also result in greater secondary wastes including those arising from decommissioning of the additional facilities. It has been assumed that Option G2 is implemented with a carbon capture technology and there could be a significant amount of waste arising from this process.

- Option G1 would not require abatement to minimise radioactive gaseous discharges to the environment whereas Option G2 would. The unabated doses due to discharges have been estimated at 2.8E-01 µSv/y for Option G1 and 5.5E+02 µSv/y for Option G2. Following abatement this dose falls to 1.3E+01 for Option G2, which is still two orders of magnitude greater than the dose for Option G1, Option G2 is also reliant on a carbon capture technique to minimise discharges of carbon to the environment, the efficacy of which is unknown. In Option G1 there would be little interaction between the waste and the environment during encapsulation or storage. Interaction with the environment following disposal through radionuclide leaching is mitigated by the encapsulated waste package concept, which is well underpinned in this respect.
- Option G2 is also more demanding on materials and resources, including a highly energy \bullet intensive process, and incurs greater off-site disturbances and an approximate doubling of waste-miles compared with Option G1 (estimated at 17,800 (G1) vs 37,020 (G2)).

Overall, option G1 performs best as, although it results in a significantly greater volume of waste being disposed of, it has an otherwise much smaller impact on the environment as it avoids the materials and energy demands of a box emptying facility and graphite crushing and gasification process. Discharges to, and interactions with, the environment are comparatively small.

4.2.3 Technical Attributes

The RAG 'score' for each option against each attribute (from Appendix E) is shown below:

The key differences between these options are:

- Option G2 applies established techniques to the point of treatment, though it is not certain that sorting and segregation could be performed to the requirements for treatment and there is an increased risk due to 'raw' storage. There would be considerable competition for shared assets/space at site due to the requirement for a 3m³ box emptying facility and use of the reactor decommissioning encapsulation plant. The treatment process applies existing technology but in a highly novel manner and there is a credible development risk associated with this. The technical cases required to support management of the treatment by-products are unknown. In comparison. Option G1 applies a well-established management strategy and uses existing plant and processes, which provide high confidence that the technical objectives can be achieved.
- The key technical risk with Option G1 regards uncertainty over demonstrating at-depth disposal WAC, including whether sufficient information exists and will be retained to support this future assessment. This risk also applies to the final management of waste in Option G2, for a smaller number of higher activity packages. The information risk can be better mitigated in Option G2 as information deficiencies, if understood, could be addressed when the waste undergoes further handling.

Overall. Option G1 performs best as it presents the least challenging route to final management. applying a well-established management strategy and using existing plant and processes. Option G2 performs less well as it introduces additional complexity and risk, particularly as a novel treatment process is used, and presents significant competition for site assets. The key risk with Option G1 is

demonstrating that at-depth disposal WAC, when they are established, can be met and that there is sufficient information about the waste to support this demonstration.

4.2.4 Socio-Economic Attributes

The RAG 'score' for each option against each attribute (from Appendix E) is shown below:

The key differences between these options are:

- Option G2 requires a new facility for emptying 3m3 boxes and sorting/exporting waste for treatment, at great cost, whereas Option G1 makes use of existing plant and equipment. Overall, Option G2 is estimated to cost double that of Option G1 (£32.5M (G1) vs £65M $(G2)$).
- Option G2 would also defer management of the waste to a future generation whereas in \bullet Option G1 the waste would be promptly encapsulated with no future processing demand and, although the waste packages will require management through to disposal (which will be performed by a future generation), this demand is significantly less than that placed on future generations by Option G2. It is also not clear what this burden entails for Option G2 as the long-term management requirements for gasification by-products are unknown.

Overall, Option G1 provides the most cost-effective route to final management and leaves no further waste processing for future generations to do. Option G2 is greatly expensive and places a large burden on future generations due to the deferred sorting and management and the long-term management requirements of the treatment by-products are unknown.

4.3 **Lead Option(s)**

The option to encapsulate then store for at-depth disposal (Option G1) is the lead option for all waste from Bunkers 2-5.

This is because the alternative option of gasification performs worse in nearly every attribute, with the key differentiators being:

- Sorting the graphite and exporting it for gasification involves additional handling steps and \blacksquare increased worker dose. There would be greater conventional safety risks from activities associated with construction, operation and decommissioning of the $3m³$ box emptying facility, and greater inherent hazards from the gasification process itself.
- Gasification would result in comparatively large radioactive and non-radioactive wastes to \bullet manage, reliant on abatement and carbon capture techniques to limit discharges to the environment. The process is also greatly energy intensive. By comparison, the at-depth disposal option would result in very little interaction with the environment.
- The burden placed on future generations would be significantly greater with the gasification option and it would place significant strain on site assets and resources, presenting a large scope of work to be conducted during the same phase as reactor decommissioning.
- Gasification technology is at a very low level of readiness and there is a credible risk that the technology will never materialise.
- Greater uncertainty with management of residual HAW which is unlikely to be suitable for \bullet near-surface disposal.

Significantly greater cost (£65.0M vs £32.5M). \bullet

Sensitivity Analysis 4.4

This section considers the sensitivity of the lead option against the relative assumptions from Appendix B, replicated below:

DETERMINATION OF THE LEAD OPTION FOR BUNKER 1 5

5.1 **Bunker 1 Waste Summary**

The inventory of waste from Bunker 1 is summarised below. Some of this waste has been retrieved and is in storage in 3m³ boxes in the ILW Store, some of this waste remains in Bunker 1. For the retrieved portion, during packaging the waste was processed (sorted, size reduced, etc.) according to the requirements of the LoC Waste Product Specification (WPrS) [14] which includes limits on certain waste items, such as fuel fragments¹¹ and HEPA filters, and identifies measures to segregate particulate waste. Where possible low level waste items are segregated for separate management.

To assess management options for waste from Bunker 1 it is necessary to consider the 'in-bunker' portion and retrieved portion separately. The potential for alternative management is greater for the in-bunker portion and so this is considered first. The implications of this assessment to the waste already retrieved from the bunker are discussed in Section 5.3.

5.2 Assessment of Options for Managing Waste from Bunker 1– In-Bunker Portion

This section presents a comparison of options for managing the in-bunker portion of waste from Bunker 1, based on the attribute comparison table in Appendix E and assumed steps involved with each of the options (set out in Appendix B).

The attributes of each option have been assessed and 'scored' relative to each other by virtue of assigning a green/amber/red colour according to the system set out in Reference [3], where:

- × Red is allocated where the performance of the option on that issue is significantly worse than at least one other option.
- Amber is allocated where the options is clearly, but not significantly, worse than at least one × other option.

 $¹¹$ Reference [14] allows for the packaging of up to 1.0kg of segregated fuel fragments in a fuel handling vessel. This</sup> strategic options assessment has excluded segregable fuel fragments and so does not wholly align with Reference [14].

Green is allocated in other cases.

The relative performance of each option is discussed within each attribute category as follows:

5.2.1 Safety Attributes

The RAG 'score' for each option against each attribute (from Appendix E) is shown below:

The key differences between these options are:

- Option M1 makes use of existing plant and processes without modification. All other options require some degree of construction/modification with associated conventional safety risks due to construction and implementation. These risks are viewed to be greatest for options requiring new facilities for 3m3 box emptying (Options M5 and M6) or dissolution (Option M3). Options based on high hazard processes (Options M4 and M6) are viewed to present greater inherent conventional safety risks. Additionally, Options M4 and M6 present greater risks due to their novelty. All waste is encapsulated in Options M1, M2 and M5 and so encapsulation hazards are not viewed to be discriminating between these options but the additional facility requirements for Options M2 and M5 are. Due to the properties of magnox there are flammable gas hazards with each option, and this assessment has not discriminated between options on this basis (though some options are likely to require less active management than others).
- In Options M1-M4 the waste would be passivated by encapsulation or treatment in the \bullet near-term (20 years) whereas Options M5-M6 defer passivation to the reactor decommissioning phase and so present a comparatively long time at increased risk due to 'raw' storage (see Section 6.3.1 for consideration of these risks).
- The least worker dose would be expected with Option M1 because all other options require some degree of additional waste handling. Whilst any option would be managed in such a way as to ensure doses are ALARP, the overall ALARP option would be one which avoids the work in the first place, i.e. Option M1.
- Options M1, M4 and M6 make the most radioactive transports (100, 226 and 241) and Options M32, M3 and M5 make the least (36, 16 and 59). Public doses are considered to be greater for the former options, though the effect is likely to be small and this assessment has not considered the shielding properties of the different transport packages. Conventional traffic risks would also be greater for Options M3-M6 which have a greater number of rad-waste consignments and/or the need to import additional materials to site for construction.

Overall, Option M1 performs best on safety as it makes use of existing plant and processes, and the waste is promptly passivated (by encapsulation). Other options perform less well as they involve

new facilities and processes being established at site, increased handling of the waste, greater number of transports and/or delays to passivation.

5.2.2 Environment. Waste and Sustainability Attributes

The RAG 'score' for each option against each attribute (from Appendix E) is shown below:

The key differences between these options are:

- Option M3 results in the smallest volume of waste for disposal / final management. Option M1 performs least well as it would dispose of all waste without first reducing its volume or activity. All other options achieve some degree of volume reduction. The activity of waste disposed of / finally managed is also reduced for Options M3, M4 and M6 due to discharges during treatment. Options involving further handling would also result in greater secondary wastes such as those arising from decommissioning of contaminated plant.
- Encapsulation options (M1, M2 and M5) have been estimated to result in doses due to \bullet radioactive gaseous discharges in the order of 10⁻³ µSv/y whereas treatment options (M3, M4 and M6) are estimated to result in doses which are two-to-three orders of magnitude greater, even after factoring abatement (see Appendix H). Non-radioactive discharges can also be expected with the treatment options. By contrast, the encapsulation options would involve little interaction between the waste and the environment during encapsulation or storage. Interaction with the environment following disposal through radionuclide leaching is mitigated by the encapsulated waste package concept, which is well underpinned in this respect.
- Options M4 and M6 apply the most energy-intensive processes. Whilst the process energy \bullet requirements are not expected to be as significant for Option M3 there would be a high demand during construction. Options M1, M2 and M5, by contrast, have a low energy demand. Material demand has not been judged to be discriminating between options; materials required for construction of new facilities are assumed to be recoverable though there would be off-site disturbances during these construction activities, most notably for Options M3, M5 and M6.
- Options M1 and M3 perform best on waste-miles, both involving on-site \bullet processing/treatment and near-site disposal/final management. All other options perform significantly worse, with Options M4 and M6 involving the greatest waste-miles. The wastemiles for each option have been estimated to range between 220 miles (M3) and 37,705 miles $(M6)$.

Overall, Options M1 and M2 are viewed to perform best and similarly, with the latter disposing of a smaller volume but incurring far more waste-miles. Option M5 (the deferred variant of Option M2)

performs less well primarily due to the disturbances associated with the construction, operation and decommissioning of the 3m³ box emptying facility.

Options M4 and M6 have been assumed to result in a similar volume reduction as with supercompaction (Options M2 and M5) but with a high energy demand and increased radiological discharges to the environment, and so is viewed to perform less well than Options M2 and M5. Option M3 would incur similar detriments but with much improved volume reductions, however, and a case could be argued for it on environment, waste and sustainability grounds if minimising the volume of waste for disposal was a priority objective.

5.2.3 Technical Attributes

The RAG 'score' for each option against each attribute (from Appendix E) is shown below:

The key differences between these options are:

- Options M1, M2, M3 and M5 apply established techniques and represent low development \bullet risks, although Option M3 could have a significant greater deployment difficulty. Options M4 and M6 apply unproven technology and so have very credible development risks though the deployment risks are borne by others (not Magnox Ltd). Options M3, M5 and M6 represent the greatest competition for shared assets at site, and Option M3 is most likely to threaten the site schedule and C&M entry date.
- Option M1 is viewed to carry the lowest risk of failure as it applies a well-established \bullet management strategy and uses existing plant and processes, though the key risk regards uncertainty over demonstrating at-depth disposal WAC, including whether sufficient information exists and will be retained to support this future assessment. This uncertainty affects all options to some degree though for 'final management'. Options M5 and M6 allow more time for WAC or guidance to be established before the waste is processed though have the greatest risks due to 'raw' storage. Options M2 and M5 have technical risks relating to demonstrating at-surface WAC, due to the heterogeneity of the waste and presence of small, high activity items and sludge contamination. It may not be possible to sort the entirety of the magnox to enable alternative management, meaning that a greater volume of waste is routed for long-term storage in Options M2, M3 and M5.

Overall, Option M1 performs best as it presents the least challenging route to final management. applying a well-established management strategy and using existing plant and processes. Other options perform less well as they introduce additional complexity and risk, particularly where novel treatment processes are used. Another key detriment is over competition for site assets and threats to C&M entry, which are greatest for Option M3. The key risk with Option M1 is demonstrating that at-depth disposal WAC, when they are established, can be met. Further information cannot be as readily gained for this option as opposed to those which involve further handling of the waste.

However, unlike all other options. Option M1 does not concentrate the activity of the waste (either by treatment or segregation) and so represents the lowest risk option for meeting these WAC.

5.2.4 Socio-Economic Attributes

The RAG 'score' for each option against each attribute (from Appendix E) is shown below:

The key differences between these options are:

- Option M1 is estimated to be least costly. Options M2 and M4 are expected to be costlier. \bullet primarily this is due to the additional infrastructure needs at site. The costliest options are those involving construction of significant infrastructure at site, i.e. Options M3, M5 and M6; these could incur 10-20 times the cost of Option M1. The cost of each option has been estimated to range between £5M (M1) and £105M (M3).
- In Option M1 the waste would be promptly encapsulated with no future processing demand. Option M2 is similar though encapsulation would be deferred slightly. Both options would require future generations to manage waste packages through to disposal and Option M2 would require that some packages be long-term stored. Options M3 and M4 would treat the waste in the near-term but require that future generations manage waste packages for long-term storage, the requirements of which are uncertain. Options M5 and M6 entirely defer management of the waste to future generations and perform least well in this respect.

Overall. Option M1 provides the most cost-effective route to final management and leaves no further waste processing for future generations to do. Options M2, M3 and M4 reduce the volume of waste for future generations to manage but leave a long-term management burden which is unknown. The greatest future burden is associated with Options M5 and M6, and these are also high-cost options due to the need to retrieve waste from 3m³ boxes. The costliest option, however, is likely to be Option $M3.$

5.3 Assessment of Options for Managing Waste from Bunker 1 - Retrieved Portion

As it is simpler to retrieve and sort the in-bunker portion of the waste than that which has already been containerised, it can be inferred that if the lead option for managing the in-bunker portion of the waste to encapsulate then store for at-depth disposal then this will also be the lead option for the retrieved portion. See Section 5.4.

5.4 **Lead Option(s)**

The option to encapsulate then store for at-depth disposal (Option M1) is the lead option for all waste from Bunker 1 (both the in-bunker and retrieved portions).

Treatment options have been assessed to perform less well compared to disposal options. This is because:

- Option M3 (dissolution) would require substantial new facilities with associated \blacksquare construction, operation, and decommissioning hazards. This would also place considerable competition for resources at site and impact the site schedule. Dissolution is a complex process with credible development risks. This option would also incur the most cost by far. estimated at £105M for the in bunker portion alone.
- Options M4 and M6 (vitrification) apply inherently high hazard and novel processes. The \bullet vitrification process has high energy demands and increased waste-miles. The technology has a low readiness level and there is a credible risk that it will never become available or proven for the waste from Bunker 1.
- With all treatment options there would be greater discharges to the environment when \bullet compared against disposal (encapsulation) options, estimated to be two-to-three orders of magnitude greater. There is also greater uncertainty with management of residual HAW which is unlikely to be suitable for near-surface disposal. There would also be significant challenges to effectively sort and segregate to the requirements of these processes.

Of the disposal options, Option M1 is judged to perform best overall. The next best disposal option is Option M2 (to sort and decay store then encapsulate for at-surface disposal). This option is, on balance, judged to perform worse than the lead option due to the following key differentiators:

- The magnox would need to be segregated from the other waste from Bunker 1 and there \bullet are risks that a portion of the magnox cannot be adequately sorted to meet at-surface disposal WAC. Challenges are presented by the heterogeneity of the waste and presence of small, high activity items, as well as particulate contamination.
- Sorting the magnox and exporting it requires additional waste handling and results in \bullet increased worker dose due to this.
- For the waste remaining in bunker the requirement for a new sorting facility would present \bullet competition for resources at site and would impact the site schedule. This issue is significantly exacerbated for the waste that has already been retrieved and containerised.
- There is greater uncertainty with management of the segregated, non-magnox waste as \bullet this is unlikely to be suitable for at-depth disposal if packaged on its own.
- At-surface disposal would also incur additional waste-miles and cost. \bullet

5.5 **Sensitivity Analysis**

This section considers the sensitivity of the lead option against the relative assumptions from Appendix B. replicated below:

MANAGEMENT ANALYSIS 6

The following sections explore the risks associated with the lead options.

Summary of lead options for consideration 6.1

The lead options are:

Both options involve the encapsulation of waste in 3m³ boxes and storage until a disposal facility exists. It is appropriate to consider them together within this management analysis.

For waste from Bunker 1 the next best option (M2, to sort and decay store then encapsulate for at surface disposal (LLWR)) is also considered to determine whether there is an overall advantage to adopting this option instead (see Section 6.3.4).

6.2 **Risks Associated with the Lead Options**

From the sensitivity analyses in the preceding sections, two key risks are highlighted for both options relating to disposal and foreclosure of options. These are considered below.

6.2.1 Disposal Risks

Disposal risks exist for all wastes which are planned to be disposed of in future facilities, whether that involves NSD or the GDF. The lead options propose that the bunkers waste be disposed of 'atdepth' in a Scottish facility vet to be designed or built. The LoC process has been followed to provide assurance that the future management of waste packages has been taken into account as an integral part of their development and manufacture [2] and a measure of disposal confidence can be taken from having secured fLoCs (as well as assuring deep geological disposal should policy change to permit this). This assurance applies not only to the waste packages but also to the quality of the information (including characterisation) held about them within the associated package records. Further assurance has been sought by assessing human intrusion scenarios for various at-depth disposal configurations (see Appendix G), which provides further confidence that the waste packages can be disposed of as intended.

6.2.2 The Risk of Foreclosing Alternative Long-Term Management Options

Scotland's HAW Policy notes that decisions need to recognise the risk of foreclosing alternative longterm management options. The lead options involve prompt packaging and conditioning by encapsulation, and this makes it difficult to change strategies away from disposal.

Should it later prove beneficial to extract materials from encapsulated waste packages it may be possible to adapt a process similar to that used for recycling concrete, though this would need further work to establish. In any case, the benefits of adopting an alternative strategy would have to outweigh the detriment of doing so.

This review has considered the possibility that alternative options become available in the future; the question is to what extent should such opportunities be enabled. Encapsulation would clearly make a strategy change more difficult than if the waste was not encapsulated. However, there is also a significant challenge associated with removing waste from 3m³ boxes, regardless of whether it is encapsulated or not. This review identifies the detriment associated with the task, highlighting that it is unlikely that an alternative strategy would be preferable; the desire to retrieve waste would have to be driven by some imperative not currently foreseen.

6.3 Risk Mitigations

6.3.1 The Potential to Reduce Disposal Uncertainty

The NDA is pursuing development of a NSD capability for England and Wales and, as set out in their strategy [15], this includes the investigation of *earlier opportunities for the implementation of nearsurface disposal solutions* in Scotland. Magnox Ltd is supporting the NDA with this work and will continue to work with the NDA, other Scottish waste producers, and regulators to further how NSD considerations factor into near-term decision making. This is not recommended as a specific action from this review as it is captured by ongoing work and is not specific to the bunkers waste.

6.3.2 The Potential to Defer or Forego Encapsulation

One mitigation against disposal risks is to defer encapsulation until a disposal facility exists. The assumption in this review is that such a facility would not arise until after the Hunterston A site has been decommissioned. There is the possibility that a disposal facility would be available earlier than envisaged, however there is a greater possibility that the reactor decommissioning phase for the site will be brought forward¹². The assumption used in this review is therefore viewed to be likely and that, if the intent was to encapsulate the waste only once a disposal facility was established, this would need to be done outside of a planned site decommissioning phase. The greatest potential to accommodate this could be if the disposal facility had an encapsulation capability, though this is not the current plan.

Encapsulation forms part of the design intent for waste stored in thin-walled containers such as the 3m³ box. During storage the grout limits waste and container degradation, assuring their ongoing performance. Following disposal the grout provides retardation of release of radionuclides into the biosphere and stabilises chemically reactive or mobile materials, as well as providing a structural function.

Reference [16] considers the risks associated with deferring encapsulation, from a waste package integrity perspective. Risks can generally be said to be greater for waste from Bunker 1 than for Bunkers 2–5, due to the greater reactivity of the magnox. Waste degradation could present challenges, particularly where corrosion products have formed which inhibit the ability for grout to infiltrate the package. Container degradation issues could also present.

It may be possible to forego encapsulation entirely. Nuclear Waste Services (NWS) waste package specification notes *the possibility to package certain types of waste [in the 3m³ box] without the use of an immobilising medium exists* [17], and Magnox Ltd has recently sought their expert view on the potential for the bunkers waste packages to be disposed of to the GDF without a grout encapsulant. This expert view is provided in Reference [18] concluding that *disposal in an unencapsulated state would be inconsistent with [NWS] requirements as currently understood for geological disposal* and highlighting several disposal risks, the most severe of which related to GDF operations safety (ALARP) and post closure structural/geological concerns (voidage). It was also noted that continued storage of the waste without encapsulation *may create an environment where the internal surfaces of the stainless-steel box are vulnerable to corrosion.* Whilst this expert view was provided with reference to the GDF these highlighted risks would apply a near-surface facility, and voidage risks are potentially exacerbated.

To defer (with the potential to forego) encapsulation goes against Magnox Ltd waste management principles, set out in Reference [19] and derived from regulatory guidance, best practice, etc. These combine to present a general imperative for prompt passivation. The significance of adopting a strategy of deferred encapsulation should not be understated. This review has not identified grounds for pursuing such an approach.

 12 Update for Issue 2 of this report: Magnox Ltd has formally changed its strategy to a rolling programme of decommissioning and physical work is now forecast to have completed at Hunterston A in the 2050s.

6.3.3 The Potential for Policy to Change

Whilst not considered as an option within this review, there is the potential for policy change to enable a GDF route for Scottish HAW, which would mitigate at-depth disposal risks and is supported by fLoCs for the bunkers waste packages.

6.3.4 Would Another Option be Preferable?

The assessments summarised in Sections 4 and 5 provide no indication that treatment options would be preferable to disposal options. The only alternative disposal options are identified for waste from Bunker 1 and the best of these was identified as Option M2: to sort and decay store then encapsulate for at surface disposal (LLWR).

In this option most of the waste (the magnox) would be decay stored pending encapsulation and disposal. The waste would be encapsulated at the disposal facility (LLWR) and so is not dependent on the availability of an encapsulation facility at the Hunterston A site. Greater flexibility is therefore provided with this option for the magnox until the point of encapsulation, which provides some mitigation against the risks considered above.

The non-magnox waste items would be packaged and promptly encapsulated and so this waste is subject to the same risks considered above, however these risks are exacerbated as although the magnox represents ca. 90% of the volume of waste from Bunker 1 it accounts for less than 3% of the total activity. The non-magnox waste is much less likely to be suitable for at-depth disposal if packaged on its own.

There are also risks with Option M2 by introducing further sorting and segregation. There is mixed experience with FED sorting at other sites. It proved possible to sort and divert a large portion of FED from Bradwell to LLWR. It was not proven possible for FED at Hinkley Point A, which is being managed as ILW and has significant challenges even sorting to IP-2 transport requirements. The strategy for managing FED from Oldbury and Sizewell A is being reviewed in part due to learning from Hinkley Point A. For the FED from Bunker 1 it would not be possible to segregate materials without cross-contamination, especially given the presence of pond sludge and graphite particulate waste, and the ability to effectively segregate small volume, high-activity MAC would also need to be tested. It is not a given that the waste could be sorted to the requirements of the LLWR WAC, and probable that a portion would need to be routed for at-depth disposal in any case.

In general terms, options M1 and M2 apply similar strategies, i.e. to package and store pending disposal, and the key differentiator is the additional sorting and segregation for Option M2. Option M2 would provide greater disposal certainty for the portion of magnox which can be effectively segregated (as this can be assessed against established WAC) however it exacerbates disposal risks for the remaining waste, whereas Option M1 appears to provide greater overall assurance that all waste can be disposed of.

From this consideration it is judged that no alternative options are preferable to those identified as the lead options, Options G1 and M1. It is recommended that these be taken forwards as the proposed options (and, as these options apply the same strategy, this can be expressed as a singular proposed option).

6.4 Contributors to this Management Analysis

Contributors are listed in Appendix A.

6.5 Proposed Option

The proposed option is to manage all waste within the scope of this review via encapsulation for interim storage then at-depth disposal.

This assessment identified alternative options for treatment or disposal of the waste from Bunkers 1-5. All treatment options were judged to perform less well than disposal options because:

- All treatment options require new facilities to process the waste, requiring additional waste handling and resulting in increases to worker dose. There would also be conventional safety risks from the construction, operation and decommissioning of these facilities, and particularly for the scenarios where waste needs retrieving from 3m³ boxes.
- Most treatment options apply inherently high hazard processes and, aside from dissolution, all technologies have low technology readiness levels and credible development risks.
- All treatment options would result in gaseous and/or liquid wastes to manage. Even following abatement it was estimated that doses due to radioactive discharges would be two or more orders of magnitude greater than those with the disposal (encapsulation) options. The estimated total doses (following abatement) due to discharges for each option¹³ are shown below in μ Sv/y:

Applying BPM, as set out in SEPA guidance [4], presents an imperative to minimise radioactive waste discharged to the environment, a necessary consequence of which is that the amount sent to solid disposal facilities is maximised.

- All treatment options introduce technical challenges to adequately sort and segregate the waste to the requirements of the relevant process.
- All treatment options result in HAW which is unlikely to be suitable for at-depth disposal.
- Some treatment options would incur significantly greater cost. The estimated total cost of each option¹³ is shown below:

Alternative disposal options were only identified for magnox from Bunker 1, however at-surface disposal performed less well than at-depth disposal because:

- The magnox would need to be segregated from the other waste from Bunker 1 and there are risks that a portion of the magnox cannot be adequately sorted to meet at-surface disposal WAC. Challenges are presented by the heterogeneity of the waste and presence of small, high activity items, as well as particulate contamination.
- Sorting the magnox and exporting it requires additional waste handling and results in increased worker dose due to this.
- For the waste remaining in bunker the requirement for a new sorting facility would present competition for resources at site and would impact the site schedule. This issue is significantly exacerbated for the waste that has already been retrieved and containerised.
- There is greater uncertainty with management of the segregated, non-magnox waste as this is unlikely to be suitable for at-depth disposal if packaged on its own.
- At-surface disposal would also incur additional waste-miles and cost.

The proposed option, to manage all waste by encapsulation for storage then at-depth disposal, follows an underpinned waste management approach and is assured via the LoC process. The key risks with this approach relate to disposability and foreclosure of options. The possibility to defer encapsulation has been considered as one tactic to mitigate these risks, however this review has identified no grounds for pursuing such an approach. Prompt encapsulation is proposed to ensure that the waste packages are manufactured in accordance with the design intent of the $3m³$ box, to

¹³ Figures presented for Bunker 1 options (M1-6) are for the 'in bunker' portion of the waste only.

provide better long-term performance during storage and during the disposal facility's operational and post closure phases.

Implementation of this strategy will continue to be managed using the Hunterston A site Radioactive Waste Management Case, which will be updated to reflect the findings of this review.

6.6 Consistency with Scotland's HAW Policy

The proposed option aligns with the baseline strategy which was recently reviewed for its consistency with Scotland's HAW Policy. This review was documented in Reference [20], concluding that the baseline strategy demonstrated good alignment. This review is supported by the analysis of *this* options assessment which has considered all long-term management options (so far as they can be currently determined) taking account of the fundamental principles set out in the Policy, the waste hierarchy, and proximity principle. The lead options satisfy the proximity principle, and alternative waste management options, which enable greater application of the waste hierarchy (through volume reduction), have been given due consideration though the detriments of these options have been judged to outweigh the benefits.

As noted in Reference [20], decisions on waste management are influenced by numerous things including policy, regulatory guidance, etc., and that these generally influence decisions in favour of prompt packaging and conditioning (passivation), to immobilise radioactivity and limit waste degradation for reasons of passive safety amongst others including intergenerational equity. The lead options determined from this review align with this generality.

7 BPM AND ALARP SUMMARY ARGUMENTS

7.1 Summary BPM Argument

This document presents a strategic options assessment. The proposed option is considered to represent BPM and supports the requirement to optimise public exposures as:

- it results in minimal generation of further radioactive wastes (secondary wastes are limited to operational and decommissioning wastes associated with existing retrieval and encapsulation processes/equipment, which has arisen/will arise in any case); and
- it minimises the potential for radioactivity to be discharged to the environment (discharges would arise primarily from encapsulation off-gassing, which would be similar in all disposal options assessed; treatment options would result in comparatively significant discharges).

The strategic assessment presented in this document has not addressed how the implementation of the proposed option will be optimised. To clarify how this is being achieved an optimisation summary report will be produced. This will set out how the radiological effects of radioactive discharges on the environment and people have been or will be minimised.

7.2 Summary ALARP Argument

The proposed option is considered to represent ALARP. No additional handling of the waste is proposed (whereas all other options would involve this) and it makes use of the site's established HAW management route which is remotely operated by and large. There would be minimal off-site doses due to discharges compared with treatment options. There are other options which would result in fewer radwaste consignments, and potentially lower public doses due to transports, though these are overall less preferable due to other factors. Long-term risks are managed through prompt waste passivation and applying an underpinned waste management approach.

8 FURTHER WORK

8.1 Further Work Required to Support the Outcome

No further work is proposed to support the outcome of this strategic options assessment. Forward actions may be needed to optimise its implementation, these will be identified upon completion of an optimisation summary report
Whilst not forming a forward action from this review, it is noted that work is ongoing to support the NDA in developing a NSD capability and that this work, along with collaborations with other Scottish waste producers, will hopefully improve how disposal uncertainties are factored into near-term decisions.

9 REVIEW REQUIREMENTS

9.1 Review Requirement (When)

This is to be periodically reviewed every five years, in accordance with company standard S-391, or sooner if prompted by a trigger (see below), until the waste is encapsulated.

Following encapsulation this should only be reviewed if prompted by one of the tiggers.

9.2 Review Requirement (Triggers)

This should be reviewed if:

- LLWR WAC change to enable disposal of the waste as packaged.
- Other alternative management routes become available which challenge the assumptions used in this review.

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APPENDIX A: OPTIONS ASSESSMENT PANEL AND STAKEHOLDER ENGAGEMENT

Options Assessment Panel

The Options Assessment Panel (OAP) included the following people:

- Oliver Smith, Senior Waste Consultant, Waste Strategy & Permissioning
- Elena Alcantara, Baseline Strategy & Integration Manager, Waste Strategy & Permissioning
- Reuben Phillips, Waste Manager, Hunterston A Site
- Graeme Busby, NRE SAWBR HAW Operations, Hunterston A Site
- David Bremner, Radiation Protection & Environment (incl. RSL Adviser), Hunterston A Site
- Tom McLaughlin, Operations Engineer, Hunterston A Site
- Cameron Robertson, Engineering Lead (Waste Programmes), Hunterston A Site
- Richard Delley, HAW Disposability Officer for Hunterston A Site, Waste Strategy & Permissioning
- Michelle Grist, Senior Waste Consultant, Waste Strategy & Permissioning
- Jack Clarke, Assistant Engineer, Waste Strategy & Permissioning

Data and underpinning were provided and/or scrutinised by:

- Joshua Weatherill, Assistant Engineer, Technical Function
- Hannah Bean, Technical Lead, Technical Function
- Bill Westall, Principal Consultant, Technical Function
- Ceri Davies, Principal Waste Consultant, Waste Strategy & Permissioning
- Joseph Stephenson, Senior Waste Consultant, Waste Strategy & Permissioning
- Darryl Smith, Senior Waste Consultant, Waste Strategy & Permissioning
- Eden Nuclear and Environment Ltd
- Andy Sealby, R&D Manager, Sellafield High Active Waste Thermal Treatment Programme
- Christopher Healey, Cost Engineer
- Michelle Grist, Senior Waste Consultant, Waste Strategy & Permissioning

The following stakeholders were also engaged throughout this review:

- Tim Bond, Waste Strategy & Permissioning Manager (principal reviewer of the Management Analysis)
- Mark Blackley, Site Director, Hunterston A Site (Alistair Walker, Acting Site Director, Hunterston A Site engaged prior to Mark's appointment)
- Alan Krailing, Head of Profession, Environment & Waste
- Paul Murray, Waste Programme Director
- Paul Winkle, Chief Operating Officer
- Pam Duerden, EHSS&Q Director
- John Grierson, Deputy Chief Operating Officer

Further review was provided by:

- Andrew Oborne, Waste Programme Chief Engineer
- Patrick Haley, Principal Consultant, Environment & Waste
- Melissa Hughes, Principal Consultant, Waste Strategy & Permissioning

APPENDIX B: OPTIONS ASSUMPTIONS AND DESCRIPTIONS

Option Constraints and Assumptions

The following constraints are identified:

- Prompt processing will take place during the present decommissioning phase (ca. 2020-30) \bullet
- Deferred processing will take place during the reactor decommissioning phase (ca. 2070- \bullet $80)^{14}$.
- For all deferred processing options, all waste will be stored containerised and 'raw' (i.e. not \mathbf{r} encapsulated) in the ILW Store until the reactor decommissioning phase. The waste that is currently in-bunker will be retrieved and containerised in line with existing practice.
- Magnox Ltd would need to provide a facility for the on-site processing (retrieval, sorting, \bullet packaging, etc.) of waste:
	- The sorting of in-bunker waste, and loading into 200 drums, can be performed in \circ the SAWB with new equipment installed above bunker: waste would be hoisted from the bunker to the sorting area in buckets
	- o The emptying of 3m³ boxes for sorting and loading into 200I drums cannot be performed in an existing facility; a new building will be required.
- Wastes packaged for at-depth disposal will be packaged in line with prevailing site HAW \bullet management strategies, i.e. encapsulation in 3m³ stainless steel boxes. Waste can be encapsulated in the present decommissioning phase using SILWE.
- Magnox Ltd would need to provide a treatment facility for the dissolution of magnox as there is no impetus (UK demand) for others to provide such a facility.
- Others will provide treatment facilities for the gasification of graphite and vitrification of \bullet magnox. These technologies are not currently available (at a commercial scale).
- Others will provide disposal facilities:
	- o The existing LLWR facility can be used for at-surface disposal. The waste would be encapsulated at this facility.
	- There is currently no at-depth disposal facility. The waste would need to be Ω encapsulated on-site before consignment to an at-depth disposal facility.
- The option of long-term storage is not considered as a primary strategy, it only applies \bullet where no other options are viewed to be feasible.
- Cross-site transportation can be achieved with existing or planned capabilities.
- Off-site transportation can be achieved with existing or planned capabilities.

The following assumptions apply to the assessment of the options:

Assumption	Reasoning
It is feasible to sort and segregate the waste into its constituent materials to enable their separate management.	It is feasible in principle but unproven. There is mixed experience with FED sorting at other sites. It proved possible to sort and divert a large portion of FED from Bradwell to LLWR. It was not proven possible for FED at Hinkley Point A, which is being managed as ILW and has significant challenges even sorting to IP-2 transport requirements. The strategy for managing FED from Oldbury and Sizewell A is being reviewed in part due to learning from Hinkley Point A. For the FED from Bunker 1 it would not be possible to segregate materials without cross-contamination, especially given the presence of pond sludge and graphite particulate waste, and the ability to effectively segregate small volume, high-activity MAC would also need to be tested.

¹⁴ It is anticipated that the dates for this phase might change, something which is considered in Section 4.4.

Options Descriptions

The assessed options for managing graphite and magnox involve variations of:

- Timing of processing: none, prompt or deferred \bullet
- Encapsulation for disposal: at-surface or at-depth
- Treatment: dissolution, gasification or vitrification

A breakdown of steps involved with each option is included towards the end of this appendix. A basic description of each of these processing, disposal and treatment variations is given below.

Timing of Processing

No Sorting

The containerised waste is encapsulated with no further sorting.

Prompt Processing

The waste is processed during the current decommissioning phase, prior to the site's entry into its quiescent phase. This requires any waste currently remaining in-bunker to be retrieved. For waste already containerised in 3m³ boxes this requires boxes to be emptied of their contents.

Deferred Processing

The waste is processed during the reactor decommissioning phase, with all waste stored in $3m³$ boxes in the ILW Store until this phase, when all boxes would be emptied of their contents. This requires any waste currently in-bunker to be retrieved and containerised in the current decommissioning phase.

Impact of Technology Availability on Graphite Management

Although graphite treatment is proven at a small scale, commercial scale treatment options for the volume of graphite from the bunkers are not currently available. The demand for graphite treatment across the nuclear decommissioning industry will be significantly greater than that from Hunterston A alone and, therefore, it is assumed that such treatment options would be developed off-site, and that storage of the retrieved Hunterston A graphite would be required until such an option is available. The question then arises as to whether it is better to empty the $3m³$ containers and promptly process the waste, while the site is currently mobilised, or to defer processing until the route is available.

Prompt processing would involve sorting the waste and loading the segregated graphite into containers suitable for storage and later export for off-site treatment. The container choice is likely to be between the 3m³ box and 200 drum: the former would enable storage within the ILW Store but has no current transport solution, whereas there would not be space within the ILW Store for the latter though it can make use of existing transport solutions and such drums are typically preferred for importing to treatment facilities.

Given that there is not space in the ILW Store to accommodate graphite in 200l drums and there is doubt as to whether sorting the waste for storage in $3m³$ boxes is worthwhile, given uncertainty over treatment facility WAC and a reliance on the SWTC, the option of prompt processing is viewed to be sub-optimal when compared to the deferred processing option. It is therefore eliminated from further assessment.

Note: as the volume of magnox is significantly less than that of graphite, there is sufficient space to store it at Hunterston store in 200 l drums pending the availability of a treatment option, and therefore both prompt and deferred processing options are assessed.

Near Surface Disposal

The Implementation Strategy for Scotland's HAW Policy [2] envisages that a near-surface disposal facility (or facilities) will be constructed in Phase 3 of the policy's implementation, after 2070, and points to the guidance on requirements for authorisation of such facilities (the GRA) [11].

At-Depth Disposal

At-depth disposal is a form of near-surface disposal, where disposal vaults / caverns are constructed several tens of metres below the surface and accessed through a drift [22].

There is no at-depth disposal option available to the UK at present, though it is being pursued by NDA as an alternative, for some waste, to deep geological disposal [23] in addition to forming a key component of Scotland's HAW Policy. The indicative suitability of the bunkers waste for atdepth disposal has been assessed using the GRA and existing studies [24] and reported on in Reference [10]. This work indicated that the waste packages arising from each bunker are likely to be disposable based on calculating the resulting doses from bounding human intrusion scenarios at various decay periods into the future, where all scenarios resulted in doses below the GRA guidance levels within the assumed period of institutional control for the disposal facility.

For at-depth disposal options, it is assumed that:

- Waste packaging will use the prevailing packaging concept at site, i.e. encapsulation within thin-walled stainless steel containers.
- Encapsulation is required prior to disposal (see Section 6.3.2 for a commentary on this) and this happens at the Hunterston A site.
- Interim storage is required prior to disposal in any scenario, for which the existing ILW Store will be used. Should the interim storage phase extend beyond the lifetime of the ILW Store then a replacement store will be constructed. This is the basis of the LTP, with replacement stores planned to achieve up to 300 years of on-site storage.
- Waste packages will be transported for disposal within the NWS SWTC.

At-Surface Disposal

At-surface disposal is another form of near-surface disposal, where the disposal facility is on or below the surface where the protective covering is of the order of a few metres thick. Waste containers are placed in constructed vaults and when full the vaults are backfilled [22].

The LLW Repository, operated by LLWR Ltd, is one such facility and is used for the basis of assessment in this review. The suitability of the bunkers waste for at-surface disposal has been assessed using the current LLWR WAC (v5.0) and reported on in Reference [12].

For at-surface disposal options, it is assumed that:

- Magnox intended for at-surface disposal, either with or without a decay storage period, would be packaged in 200l drums and transported to a third-party facility for super-compaction, decay storage (if applicable) and consignment for disposal. This process follows that described in Reference [25]. The third-party facility is assumed to be Tradebe's Winfrith site, as this is where the existing capability has been established, though a site located closer to the Hunterston A site could be sought if pursuing this option.
- Encapsulation is required prior to disposal (see the Management Analysis for a commentary on this) and this happens at the disposal site (LLWR).

Treatment

Treatment options for managing the various waste materials have been investigated, as summarised below. With the view that other treatment options, to those listed below, may also become available in future, this review has considered a specific 'future' treatment option for both graphite and magnox which can serve as proxies for considering the benefits and detriments of future options in general.

It is assumed that any HAW by-products will be loaded into 3m³ boxes and returned to the Hunterston A site for processing (if required) followed by long-term storage.

Graphite - Gasification

There has been lots of research conducted into treatment options for graphite and several means of decontamination are theoretically possible [26]. Options for recycling or reuse could exist, e.g. 'diamond batteries', though are currently subject to great uncertainty. Gasification has been used for basis of assessment in this review as it has momentum in the industry and may be the most likely option to materialise.

Gasification involves the conversion of carbonaceous materials, such as graphite, into carbon monoxide and hydrogen by reacting the materials at high temperatures with a controlled amount of oxygen or steam, and then converting the gases to carbon dioxide and water (steam). The overall chemical reaction involves the oxidation of carbon by oxygen to form carbon dioxide.

Significant volume reductions are possible from this process, as evidenced through trials [27], though it is not clear what could be expected on a commercial scale. Secondary wastes are expected in the form of residues containing non-volatile radionuclides, the radioactivity having been concentrated. Volatile radionuclides would require management; several options exist such as gaseous discharge to atmosphere, (carbon) capture, or the formation of stable solids.

For this assessment, it is assumed that gasification results in a residue equivalent to 5% of the original volume of graphite (a reasonable guess made from trial data indicating that a significant amount of the graphite could be oxidised by this process [27] but with little data concerning its other by-products) and that this residue contains radioactivity in a concentrated form which renders it unsuitable for near-surface disposal.

Magnox - Dissolution

Dissolution has been applied at Dungeness A and Bradwell, and the technologies applied at either site could be replicated at Hunterston A. The process involves dissolving magnox in acid. Nonsoluble items will remain following the process, and by-products will be generated typically in the form of sludge, ion exchange material and filter materials. For this assessment it is assumed that dissolution results in a volume of secondary wastes totalling 5% of the original volume of magnox following encapsulation (i.e. equivalent $3m³$ box numbers are reduced by 20:1), based on secondary waste arisings from the Bradwell plant. The majority of radioactivity would be retained in these secondary wastes though a significant amount, in relation to the site's annual limit, would be discharged to environment (aqueous).

Magnox - Vitrification

This treatment technology would convert the waste, with the addition of silica-based material, into a stable glass product through the process of vitrification. It is possible that the magnox could be managed this way using a thermal treatment capability being developed at Sellafield. Further work would be required to establish the feasibility of this route along with processing constraints, should the option appear attractive. The resulting product would be suited to disposal, in general, though the process would concentrate the radioactivity of the waste and potentially prevent it from being disposable in the near-surface environment.

The plant is understood to accept waste in 200l drums and, for magnox, would most likely involve a plasma lost crucible process similar to that being trialled for drums of plutonium contaminated material (PCM). The waste would be loaded into a sacrificial container (crucible) and vitrified using a plasma torch. The crucible and wasteform could then be loaded into a $3m³$ box [28].

ISO VIEW - LOST CRUCIBLE REFRACTORY IN 3m3 BOX

For this assessment it is assumed that these waste packages would be transported to Hunterston A but require no further processing prior to entering long-term storage.

A significant advantage of this process is that it requires little sorting and segregation. For this assessment it is assumed that the degree of in-bunker sorting currently performed at Hunterston A would be sufficient, minimising requirements for additional equipment in the SAWB, though a 200l export capability will be needed. However, if transporting 'all' bunkers waste in 200l drums it is unlikely that FHISOs would be sufficient. For this assessment the Type B Novapak is assumed.

The volume reduction of the bunkers waste could be quite significant. Combined with the need for glass-forming additives and the lost crucible, the process is expected to result in a maximum of 50 200 drums of waste being processed into a $3m³$ box [29], and this upper estimate is used for the basis of assessment in this review. This is likely to be very optimistic (as a blend of materials would be needed for vitrification the amount of magnox within each box would be limited; if the waste from Bunker 1 cannot be blended with waste from elsewhere, e.g. due to policy complications, there could also arise a need for a large amount of 'clean' feedstock materials). Should this option evaluate favourably/marginally this assumption will need revisiting.

This technology may become available in the 2020s though uncertainty with this date and what the 'order book' would look like suggests that, given the plan for Hunterston A to enter C&M in 2030, some storage at site could be required prior to consignment. Storage in stillages in the ILW Store is assumed with consignment to Sellafield in the 2030s.

Other materials

No treatment routes have been identified as viable, due to activity (WAC) issues and/or proportionality (waste volume) issues. For example, it was not viewed to be credible to pursue the decontamination of metallic materials for recycling. There is the opportunity for some materials to be managed via the above routes, in particular using vitrification, though only where this would be the strategy for the dominant waste material (graphite or magnox).

Processing Steps for Management Options for Bunkers 2 - 5

The following table outlines the steps assumed within each management option. Appendix B provides an explanation of why there is no prompt processing option being assessed.

(storage phases are shown in italics for ease of identification)

Processing Steps for Management Options for Bunker 1

In-Bunker Waste

The following table outlines the steps assumed within each management option, for the in-bunker waste only.

(storage phases are shown in italics for ease of identification)

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¹⁵ Assumed for this assessment that the existing containerisation facility which is part of the Bunker Retrievals plant is no longer available or unsuitable, but an opportunity would exist to reuse this.

Containerised Waste

(storage phases are shown in italics for ease of identification)

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¹⁶ Encapsulation in reused 3m³ packages using existing site plant currently assumed

APPENDIX C: SELECTION OF ATTRIBUTES

Safety Attributes

Environment, Waste and Sustainability Attributes

Technical Attributes

Socio-Economic Attributes

APPENDIX D: ASSESSMENT DATA

The following data have been used for this assessment and originate from a variety of sources, which are quoted where relevant, and estimates (see project file for cost breakdown).

Waste volume and container loadings

Option parameters

¹⁷ Figures proportioned from total volume, 100/210

¹⁸ SAWB retrieval project roughly at 110 boxes containerised at May 2021

¹⁹ This loading is based on a 100% fill and acknowledged to be highly optimistic

²⁰ 70 drums per FHISO

²¹ 140 compacted drums per HHISO

 22 Although the waste volume reduction is much greater than this, the lost crucible process appears to result in an approximately 3:1 reduction in 3m3 box numbers (based on optimistic projections).

Option-specific container estimates

Management Options for Bunkers 2-5

Management Options for Bunker 1 – in-bunker waste only

Management Options for Bunker 1 – in-bunker waste only

Not calculated (see Section 5.3).

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²³ 802 empty boxes with assumed loading of 20 size-reduced boxes per HHISO (~40) (this loading appears reasonable, based on a ~32,000kg HHISO load capacity, ~900kg mass of each box, and accounting for furniture within the HHISO).

²⁴ 8 2001 drums per Novapak transport assumed (comprising two Novapak transport packages)

 25 100 empty boxes with assumed loading of 20 size-reduced boxes per HHISO (5)

Third Party Service Cost Assumptions

Facility Cost and Schedule Assumptions

Container Costs

Container	Purchase price
$3m3$ box	£13,946
HHISO	£13,300
200 drum	£51
200l drum stillage	£1,900

²⁶ Based on GDF disposal costs for 3m3 boxes, from NDA cost calculator v.4.2.

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²⁷ Based on total management fees for drums of magnox from Bradwell (which includes super-compaction, storage, HHISO packaging and transport but not disposal costs).

²⁸ Based on quote for incineration of magnox $(£4.1M$ for 5,579 drums). Probably optimistic.

Assumptions for Discharge Assessments

* Assessments made for in-bunker portion only (see Section 5.3).

APPENDIX E: ASSESSMENT AGAINST THE SELECTED ATTRIBUTES

The following scheme has been used to assess the options against each of the attributes selected in Appendix C:

- Red is allocated where the performance of the option on that issue is significantly worse than at least one other option.
- Amber is allocated where the options is clearly, but not significantly, worse than at least one other option. ×
- Green is allocated in other cases. \mathbf{L}

Assessment of Management Options for Bunkers 2-5

The following table compares the attributes between each management option given the assumed steps shown in Appendix B:

Assessment of Management Options for Bunker 1 (In-Bunker Waste Only)

The following table compares the attributes between each management option given the assumed steps. The data for this table assumes that 110 boxes have been retrieved, with 100 'boxes worth' of waste remaining in the bunker; the data have been scaled accordingly from the bunker totals. To clarify: these numbers only account for the in-bunker portion of the waste.

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APPENDIX F: LAW DIVERSION FEASIBILITY STUDY

To assess whether it would be feasible to pursue LAW management of the bunkers waste packages or waste materials calculations were performed considering basic activity constraints of the current LLW Repository WAC. These are reported in Reference [12].

Approach Taken

This assessment used the bunkers waste inventories [9] decayed to set times (2040, 2120, 2220) and 2320) and considering two questions - whether the waste packages could be suitable for LAW management (and, if so, when), or whether any of the individual waste materials could be suitable for LAW management (and, if so, when). These questions were answered primarily with reference to specific activity levels with some commentary on other challenges, for example the acceptability of the waste package form or individual radionuclide constraints.

Results

The results from Reference [12] are shown below. The green shading denotes an activity $<$ 12 GBq/te and a red shading where $>$ 12GBq/te. For the waste materials/items an orange shading is also used to indicate where an activity could fall < 12GBq/te after encapsulation (see Interpretation section for a commentary on this).

Waste package from	2040	2120	2220	2320
Bunker 1	33.8	15.7	8.0	4.5
Bunker 2	352.1	196.5	108.4	64.2
Bunker 3	205.8	98.3	57.9	37.8
Bunker 4	165.0	51.6	33.4	24.5
Bunker 5	228.0	89.0	52.0	33.6

Average 3m³ box waste package activity (when encapsulated) – mixed waste (GBq/te)

Weighted average activity per waste material/item (GBq/te)

Interpretation

These results have considered specific activities against the basic 12GBq/te activity criterion. From this it is indicated that the magnox could be managed as LAW but for all other waste materials or waste packages the results need to be viewed with some caution. LLW Repository disposal has formed the basis of these considerations.

Waste packages

Waste packages from Bunker 1 could potentially be eligible for LLW Repository disposal sometime next century (estimated ca. 2170) based on specific activity and presuming such a facility still exists (which it is not planned to be). If this facility had similar WAC to the current one it could prove challenging to gain acceptance due to the waste package format $(3m³$ stainless steel boxes) and due to limits on discrete items. Changes to the LLW Repository environmental safety case (ESC) and, resultantly, the WAC could present an opportunity for earlier disposal (if the WAC become less restrictive, which may not be the case). In general, the potential to dispose of Bunker 1 waste packages to the LLW Repository looks highly unlikely, due principally to the timeframe needed for radioactive decay but also WAC-related challenges.

Waste materials/items

These results arise from consideration of waste which have been segregated into discrete materials/items, e.g. magnox, fuel support members, etc. A common caution to apply to these results, therefore, concerns the practicalities of segregation and the effect that contamination, e.g. from sludge or fuel particles, could have on the specific activity of any consignment intended for the LLW Repository.

The magnox can be seen to have the greatest potential to meet the disposal WAC. These results indicate that disposal in 2040 is possible (and that the magnox would, in its raw form, be suitable in ca. 2035). Given a dilution factor from encapsulation disposal 'today' may be possible, though this review has considered a decay storage period in part to account for activity variation throughout the magnox population as well as specific activity increases from other waste items / contamination – this is consistent with Reference [25] where the requirement for decay storage is in part driven by high dose items that are difficult to segregate. Additionally, experience with Bradwell FED shows that a storage period of some duration is likely to be required to manage hydrogen generation rates for transport. In general, however, it is viewed as credible to dispose of magnox at the LLW Repository in the near-term and under the current WAC though the ability to do this for all of the magnox will be limited due to sorting and segregation / WAC demonstration challenges.

The credibility of disposing of any other waste materials/items, however, has not been demonstrated from this exercise. In most cases this is simply due to specific activities being (well) in excess of 12GBq/te. For graphite and sludge there appears to be the potential for these specific activities to be brought below 12GBq/te once the dilution effect of encapsulation is accounted for. However, sludge is effectively present within the bunkers waste as a contaminant and cannot be readily segregated. Graphite is, at face value, possible to dispose of in an encapsulated form given ca. 100 years of decay storage. The key challenge to disposing of the graphite is its C-14 content which would take up a significant portion of the capacity afforded within the LLW Repository's current ESC. It is possible that ESC revisions would lead to increased capacity for C-14 and, potentially, risk-based disposals. However, it is not currently permitted to dispose of this graphite at the LLW Repository and the time until its specific activity (when encapsulated) would fall below 12GBq/te also puts it at the extremes of credibility regardless of the C-14 issue, which is significant in itself.

APPENDIX G: AT-DEPTH DISPOSAL FEASIBILITY STUDY

To assess whether it would be feasible to dispose of the bunkers waste packages, Eden Nuclear and Environment Ltd ('Eden') was contracted by Magnox Ltd to determine whether a hypothetical at depth disposal facility, containing the bunkers waste packages, would comply with the environmental regulators' Guidance on Requirements for Authorisation for Near-Surface disposal facilities on land for solid radioactive wastes (the 'GRA').

This appendix summarises the assessment reported in Reference [10] which is based on the waste package inventories used for this strategic options review. A separate assessment has also been performed to consider the LoC inventories and to investigate the effect of differing quantities of fuel particles/fragments on the outcome, this is reported in Reference [31].

Approach Taken

As reported in Reference [10], no standardised approach exists for conducting such assessments and, as such, Eden has developed an approach based on existing regulations, guidance and previous studies (precedent) from which analogies can be drawn.

Magnox has provided Eden with waste package inventories [9], tasking Eden to determine, for a variety of cases, the time in the future at which doses from human intrusion scenarios²⁹ would fall beneath the dose guidance levels of the GRA. These calculations have been performed for several scenarios, considering intrusion into:

- waste packages from Bunker 1^{30} ;
- waste packages from Bunker 2;
- waste packages from Bunker 3;
- waste packages from Bunker 4;
- waste packages from Bunker 5;
- waste packages from Bunkers 1 to 5 (five-high stack);
- waste packages from Bunkers 2 to 5 (four-high stack)
- highest dose waste packages stacked five-high; and
- waste packages from Bunkers 1 to 5 (seven-high stack³¹).

To assess these cases, the inventories have been decayed to set assessment times (50, 100, 150, 200, 300, 400, 600, 800, and 1,000 years from now). The doses arising from intrusion have been calculated for each of these times to determine when these doses fall below the GRA dose guidance levels.

The waste packages are assumed to be disposed of in a silo-type facility, which will be sited greater than 30m below ground level. Eden has chosen a facility concept similar to the disposal silo concept being considered as part of NDA-led work on near-surface disposal opportunities. Different stacking configurations have been considered. Figure 3 shows the packages from Bunkers 1 to 5 stacked five-high, where each layer is comprised of waste packages from a different bunker.

²⁹ Human intrusion scenarios are considered to be bounding for the purposes of this assessment.

³⁰ An additional case including a 1kg fuel fragment was also considered though the results of this have been omitted from this review to align with its scope. Please refer to Reference [10] for these results.

 31 A seven-high stack is the maximum design withstand for the $3m³$ box.

Figure 3: Assumed Disposal Configuration (five-high)

Doses arising from human intrusion are considered for the following scenarios³² and exposed groups, which were defined by the Health Protection Agency (now Public Health England³³) as part of a near-surface disposal study commissioned by SEPA [24]:

- Borehole drilling
	- o Geotechnical worker handling cores on-site
	- \circ Resident (Adult, Child, Infant) living on contaminated land and consuming fruit and vegetables grown on contaminated land
	- o Consumer of contaminated drinking water (Adult, Child, Infant)
- Tunnelling
	- o Tunnel worker
	- o Spoil truck driver
- Controlled intrusion
	- \circ Controlled intrusion worker clearing the area and studying excavated material

Eden has collated and applied exposure pathway parameters from previous studies, and applied several assumptions to determine the doses received by each exposed group in each scenario. A third-party, Quintessa Ltd ('Quintessa'), has conducted a peer review of this approach to confirm that it has applied the correct regulations and guidance, draws on appropriate studies and uses appropriate assumptions.

 32 Other scenarios have been excluded from consideration for the bunkers waste, as they are either not applicable or not bounding, as noted in Reference [10].

³³ Or potentially the National Institute for Health Protection by the time of this note's issue.

Results

Whilst the calculated intrusion doses do not provide a definitive answer regarding disposability, they do provide a measure of feasibility and confidence should at depth disposal be pursued.

The time relates to the period of institutional control necessary for the disposal facility post-closure - such controls are assumed to prevent human intrusion - which is assumed to last for up to 300 vears.

The results from these calculations (reported in Reference [10]) are shown below. For each table the assessment results are shown at different times of intrusions. The dose for workers is calculated for a single intrusion event, the dose for residents and drinking water consumers represents the dose in one year. The green shading denotes a dose < 3 mSv, orange shading a dose between 3 mSv and 20 mSv and red shading denotes a dose above 20 mSv.

As a rule of thumb interpretation, disposal feasibility can be viewed to have been indicated if no doses exceed the dose guidance levels at 300 years. This is true for all scenarios considered except one.

Doses from intrusion into a package from Bunker 1

Doses from intrusion into a package from Bunker 2

Doses from intrusion into a package from Bunker 3

Doses from intrusion into a package from Bunker 4

Doses from intrusion into a package from Bunker 5

Doses from intrusion into packages from Bunkers 1 to 5

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Dose from intrusion into packages from Bunkers 2 to 5

Doses from intrusion into a five-high stack of packages from Bunker 1

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The table below contains the assessment results at key times for intrusion into a seven-high stack with one waste package from each of Bunkers 1, 2 and 5, and two waste packages from Bunkers 3 and 4. Two waste packages from Bunkers 3 and 4 have been included in this stack because these two bunkers have the highest dose waste packages from Bunkers 2, 3 and 4, and Bunkers 2, 3 and 4 contain more waste packages than Bunkers 1 and 5.

Doses from intrusion into a seven-high stack of waste packages

Interpretation

These results provide some measure of confidence that it would be possible to dispose of the bunkers waste packages in an at depth facility. They support the assumption used in this review that the bunkers wastes could be disposed of.

These results also illustrate some sensitivity to how the packages are emplaced within the facility. Further sensitivities include the assumptions used for these calculations as well as waste inventory uncertainties. In lieu of at-depth disposal WAC the LoC process has been used to guide waste package manufacture which extends to the packaging specification and adequacy of supporting information such as characterisation.

Reference [31] summarises a similar assessment performed using the LoC inventories. This also supports the assumption that the bunkers wastes could be disposed of although some scenarios require longer for doses to fall below the guidance levels.

APPENDIX H: ASSESSMENT OF DOSES DUE TO DISCHARGES

This appendix summarises the results of the discharge assessment reported in Reference [32] which is based on the Environment Agency's Initial Radiological Assessment Methodology [33], waste package inventories used for this strategic options review [9] and the assumptions shown in Appendix D.

This assessment applies the same release fractions as reported in the Hunterston A site SILWE discharge assessment [34] which was updated in 2023 to incorporate more empirical data from encapsulation of FED at Trawsfynydd.

In summary (adapted from Reference [32]):

- Doses due to discharges have been estimated for the encapsulation and/or treatment processes used in each option considered in this strategic options assessment. Doses have been calculated for all Bunkers 2-5 waste and the waste remaining in Bunker 1.
- The dominant nuclide in determining dose in all options is C-14.
- Doses due to discharges are manageable in all scenarios and abatement assumptions have been factored into the assessment where abatement is viewed as likely to be required or BPM. Note, however, it is not in the remit of this assessment to determine BPM, these assumptions have been made to support a strategic options assessment only.
- Gasification exceeds the threshold for requiring abatement. Greater abatement than has been assumed for the assessment may, in reality, be likely to enable other waste to be processed at the plant during the year.
- Vitrification does not exceed the threshold for requiring abatement, however abatement is likely to be BPM and has been factored into the assessment.
- Dissolution does not exceed the threshold for further abatement. However, it was noted that almost the entirety of the dose is due to the discharge of C-14 which was not affected by the abatement technology used at Bradwell and therefore this style of abatement is unlikely to be BPM.
- In line with the findings of Reference [34], further gaseous discharge abatement is not required or viewed to be BPM for encapsulation-based options using SILWE.

The estimated doses due to discharges are shown overleaf, these doses have been determined following abatement where applied (see Ref [32]). N.B. Doses prior to abatement are shown in Reference [32] but omitted here for simplicity.

The sum of total doses for each process in each option give the following total dose in μ Sv/y for each option (as reported in the main body of this report):

Doses from processes within each option (with abatement, where applied (see Ref [32]))

* In bunker portion only

APPENDIX I: FURTHER SENSITIVITY ANALYSIS

The following sections provide a further sensitivity analysis supporting the lead options (G1 and M1) with reference to alternative options as identified below.

Work planning

It is expected (pending Government approval) that the reactor decommissioning phase will be brought forwards for the Hunterston A site. This has no bearing on the lead options.

It would reduce the likelihood that treatment options are available by the reactor decommissioning phase (for gasification in particular). For deferred sorting options the time at risk due to 'raw' storage would be reduced and time to passivation could be shortened, though the detriment with 3m³ box emptying etc. remain. If this change removes the C&M period for site then this could help reduce the competition for site assets, particularly for the dissolution option (though this would not compensate for its other detriments).

It has been decided that the Magnox Ltd scope will be expanded to include decommissioning and waste management of the Hunterston B site, in which case there are potentially other times at which the wider site would be mobilised meaning that there could be further options for when to defer retrieval and sorting to. However, the detriments associated with the deferred sorting options are largely independent of time, relating more to the scope.

At-depth disposal

Encapsulation can be performed at the disposal facility

As captured within the options assessment, encapsulation is performed at and at-surface disposal facility (based on LLWR) but is assumed not to be performed at an at-depth disposal facility. If this change in assumption was to be agreed soon this could avoid the need for encapsulation at the site (for all options). It would increase 'raw' storage risks for options which would otherwise have encapsulated waste prior to on-site storage, putting them on-par with the deferred sorting options in this respect. It would significantly reduce the scope of work performed at Hunterston A but does not remove the scope, therefore for the prompt encapsulation options it increases the burden placed on future generations. The greatest benefit of this change in assumption could be as a disposal risk mitigation (see Section 6.3).

Disposal facility becomes available before reactor decommissioning phase

This would avoid the need for interim storage beyond the reactor decommissioning phase and would be positive for Option G1/M1 in this regard. Availability of WAC prior to this phase would also be of benefit when developing plans for Options G2, M5 and M6, though the assumption used in this assessment is that wastes remaining following processing or treatment are unsuitable for disposal.

Waste remaining following processing or treatment is disposable

This would benefit Options G2 and M2-6 by providing greater certainty over the end point for waste currently assumed for long-term storage, though the difficulty is in proving this assumption 'today' given the absence of disposal WAC and disposal feasibility studies providing no clear evidence to support a change of assumption. If this could be established then it would also reduce uncertainty with the lead options.

At-surface disposal

Disposal of magnox becomes possible without decay storage

If it was possible to at-surface dispose of the magnox (to LLWR) without decay storage this would have benefits in reducing the storage burden at Hunterston A and forgo the need to manufacture stillages and enable the ILW Store to accommodate 200l drums. It would enable earlier disposal of the magnox, however the management uncertainty for the non-magnox waste would remain. The key difference between the at-surface and at-depth disposal options assessed is the sorting and segregation required to pursue the former, which remains irrespective of when the waste is disposed.

Disposal of graphite becomes possible

It is possible that the LLWR WAC could change in future to enable disposal of graphite. However, as the graphite is already containerised, the significant detriment associated with its retrieval and sorting remains.

Disposal of the waste as packaged becomes possible

It is conceivable that changes to the LLWR WAC could enable disposal of the waste as packaged in 3m³ boxes. The waste, particularly that from Bunker 1, could comply with the facility's environmental safety case if risk-based disposals became possible. There would be a number of challenges to overcome should this become a possibility, e.g. relating to discrete items criteria, however the lead options do not foreclose this opportunity.

Disposal of magnox is not possible

It may not be possible to send all magnox for at-surface disposal, e.g. due to pond sludge contamination. It may be possible to 'clean' the magnox, though this could require additional infrastructure to that considered in the assessment. More probable is that, were this option to be chosen, it would need to be accepted that the LLWR WAC cannot be demonstrated for a portion of the eligible magnox, and this would need to be routed via the long-term storage route along with the other waste materials in the bunker.

Treatment options

Treatment options do not materialise

This is a credible risk for vitrification and gasification options and this risk favours the lead options.

Treatment processes can import 3m³ boxes

This has the potential to avoid the need for the box emptying facility, though it depends on the sorting requirements for the facility. Potentially, gasification and vitrification would be quite accommodating with their WAC and so this option has some credibility. If it were the case then the containerised waste could be consigned for processing simply, e.g. in SWTCs. It is still unlikely to be attractive to pursue vitrification for magnox as the benefit from volume reduction is not great. Other detriments with the dissolution option are also viewed to outweigh this benefit. It may be attractive for the gasification option and potentially the by-products could go on for vitrification, resulting in a greatly

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reduced volume but greatly increased activity wasteform. These options would still result in discharges and other environmental detriment, e.g. due to waste-miles, as well as present socioeconomic detriments, however. As noted above, applying BPM, as set out in SEPA guidance [4], presents an imperative to minimise radioactive waste discharged to the environment.

Other options become available

The merit of such options can be considered by using the deferred retrieval options as proxies, e.g. what benefit would such an option need to provide to compensate for the detriment associated with the process of emptying $3m³$ boxes, etc. No feasible material recovery options have been identified for this review though should these become available then a greater detriment could potentially be tolerated. This is considered further in Section 6.

Transport

Unable to transport waste in FHISOs

This would have no bearing on the lead option. It is unlikely to affect magnox though is credible for graphite, for which it may be necessary to transport in a different configuration; most likely this would be reduced quantity transports at increased cost/detriment.

Loading efficiency of 200l drums is reduced to that assumed

A highly optimistic loading assumption has been used for calculations in this assessment (100%) and, in reality, this would be lower. However, this has no bearing on the lead options; it would increase drum and FHISO numbers, and consequently waste-miles, as well as increasing other detriments such as time and cost for options involving 200l drums.

Long-term storage

This has no bearing on the lead option unless at-depth disposal WAC cannot be demonstrated (a disposal risk, see Section 6).

It has been assumed that waste remaining following processing or treatment is unsuitable for nearsurface disposal. The implementation strategy for Scotland's HAW Policy recognises that there will be such wastes, referring to them as "challenging", and that work will be needed to identify management options for them. This assessment assumes that such wastes would require 'long-term storage' whilst options are developed, potentially up to 300 years. Although the Magnox baseline allows for such storage, for wastes that are expected to be disposable there is greater certainty in the end point (and it is assumed that the disposal facility would be available in the relative nearfuture³⁴). The uncertainty associated with long-term storage affects assessment of the attributes, e.g. it is unclear what burden is being placed on future generations.

Regarding waste remaining following treatment only, it may not prove necessary to return such wastes to Scotland for long-term storage, as Clause 2.04.13 of Scotland's HAW Policy [6] infers; the presumption is only that if this adds materially to the waste needing to be disposed of in the country of destination should it be returned to Scotland, and a case could reasonably be made on these grounds. However, Clause 2.04.12 notes that the export of waste from Scotland for treatment would only be permitted where it would result in the recovery of reusable materials or the treatment would make the subsequent storage or disposal of the waste more manageable. The treatment options assessed in this review could be viewed to improve the subsequent management of the waste if resulting in volume reductions, though this argument is weakened if these wastes are unsuitable for near-surface disposal. In any case, the legality of this would need testing to fully establish the feasibility of any HAW management option conducted in another part of the UK. N.B. This is not viewed to affect the at-surface disposal option as this would be a LLW route.

Should it prove feasible, and the treatment by-products not need returning to Scotland, then it removes the uncertainty associated with long-term storage and the associated detriment identified

³⁴ The NDA is also engaging with Scottish Government about potentially implementing an at-surface disposal solution earlier than indicated in the HAW Policy's implementation strategy.

in the options assessment. This makes the treatment options more attractive but this alone is not enough to overcome the other detriments with such options, which are largely associated with the increases to dose, safety hazards, discharges, costs and burden associated with (enabling) treatment rather than managing by-products.